



Modeling distribution of vegetation types in arid and semiarid rangelands of Iran through binary logistic regression and canonical correspondence analysis techniques

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Abstract

The distribution of nine vegetation types in arid and semiarid Nodushan rangelands was predicted based on the indicator species by using binary logistic regression (BLR) and canonical correspondence analysis (CCA) techniques. Nine soil variables (salinity, pH, C/N, available water, gravel, texture, gypsum, organic matter and lime) in two depths (0-20 and 20-60 cm) and three topographic variables (elevation, slope and aspect) were used for modeling. The habitat suitability was determined using the maximum sensitivity plus specificity threshold. Results indicated that the indicator species in each vegetation type was effective and efficient for modeling the distribution of the vegetation type. BLR models provided more accurate predictions than CCA models in most of the vegetation types. The vegetation type whose distribution was well modeled by CCA was also well modeled by BLR but not vice-versa. None of the techniques provided accurate results for the vegetation type that was under grazing disturbance.

Keywords: Binary logistic regression, Canonical correspondence analysis, Indicator species, Vegetation types

Introduction

Modeling the distribution of species and communities is of high interest due to a number of reasons like for analyzing the relationships between environment and species attributes such as species distribution (Guisan *et al.*, 2002) and composition (Ohmann and Gregory, 2002), identifying the potential of an area for supporting a particular species or a high level of biodiversity, successfully implementing restoration measures and to predict the impact of climate change and land use change on the distribution and diversity of species (Riordan and Rundel, 2014). A variety of statistical models have been used to relate the environmental factors to distribution of plant species (Guisan *et al.*, 1999; Chahouki and Ahvazi, 2012; Sahragard *et al.*, 2015; Zare

chahouki and Esfanjani, 2015), plant communities (Zimmermann and Kienast, 1999; Chahouki *et al.*, 2010), biodiversity (Rodriguez-Castaneda *et al.*, 2012) and life zones (Mousaei Sanjerehei, 2014). Canonical correspondence analysis (CCA) and binary logistic regression (BLR) are the most efficient used methods for predicting the distribution, composition and structure of species and communities. CCA is unique among the ordination methods in that the ordination of the main matrix representing species data is constrained by a multiple regression on environmental variables included in the second matrix. CCA is appropriate for data set with many zeros (i.e. presence-absence data) (Guisan *et al.*, 1999; Aspinall, 2002; Calef *et al.*, 2005; Chahouki *et al.*, 2012). The distribution of species can be modeled using a regression equation made with the canonical coefficients of the environmental predictors. To predict the potential distribution of the species, Euclidean distance to the centroid (weighted average) of each species in canonical space is calculated. The centroid of a species indicates the position of the species distribution along an environmental variable (Ter Braak, 1987). For presence-absence data, the centroid is simply the mean of the ordination axis values over the sampling points selected for each plant species (Ter Braak and Looman, 1986). Binary logistic regression uses maximum likelihood to estimate the probability of a dichotomous outcome. The probability can take any value between 0 and 1 (Guisan *et al.*, 2002). Most parts of arid and semiarid rangelands of Iran are covered by *Artemisia sieberi* due to the adaptation of this species to a wide range of environmental conditions (Mousaei Sanjerehei *et al.*, 2011; Hosseini *et al.*, 2013). Therefore, for modeling the distribution of the study vegetation types whose first dominant species was *A. sieberi* (Fig. 1), the second dominant species was selected as the indicator species representing the local edaphic, topographic and climatic conditions. The objectives of this study were to determine the most important environmental drivers of the distribution of 9 vegetation types in Nodushan arid

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and semiarid rangelands, to map the potential distribution of the vegetation types using CCA and BLR and to compare the efficiency of the two models for predicting the vegetation distribution.

Materials and Methods

Study area: Three topographic (elevation, slope and aspect) and nine edaphic factors (soil texture, organic carbon, pH, EC, lime, gypsum, nitrogen, available water and gravels) were used to model the distribution of 9 vegetation types (*A. sieberi*-*S. tomentosa*, *A. sieberi*-*E. strobilacea*, *A. aphylla*-*A. sieberi*, *A. sieberi*-*S. arbusculiformis*, *A. sieberi*-*P. harmala*, *A. sieberi*-*Z. eurypterum*, *A. sieberi*-*E. ceratoides*, *A. sieberi*-*I. songarica* and *A. aucheri*) in Nodushan rangelands of Yazd, Iran, using BLR and CCA. Arid and semiarid rangelands of Nodushan are located in the northwest of Yazd province in central Iran (31°46'2" N, 52°24'2" E to 32°15'2" N, 53°47'2" E) at elevations between 1540 m and 3250 m. The mean annual precipitation ranges from 75 to 170 mm. The mean annual temperature is 12.6 °C (7.3-16.9 °C).

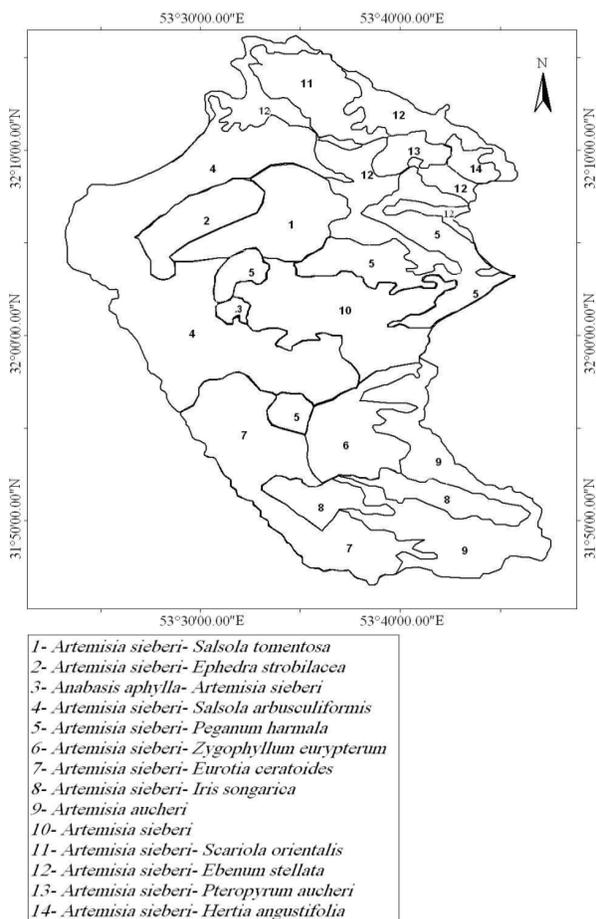


Fig 1. The actual vegetation map of Nodushan rangelands

Sampling methods: Sampling was performed in homogenous units identified by overlapping the elevation, geologic and the ETM+ cluster map (Fig. 2) of the study area. To produce the cluster map, a scene of landsat 7 ETM+ image (Path: 162, Row: 038) covering Nodushan rangelands was acquired under clear atmospheric conditions on August 06, 2006. Clustering (unsupervised classification) was performed using ETM+ bands 3 (red) and 4 (visible near infrared) images. The images of these bands effectively represent the reflectance attributes of the soil and vegetation in arid environments (Mousaei Sanjerehei, 2014). In each unit, six to ten 100-m transects were located approximately equidistantly. Soil samples were taken in the middle of each transect from the depth of 0-20 and 20-60 cm (Brunetto *et al.*, 2011; Ghorbani *et al.*, 2015). The percentage cover of each species was estimated in 30 equidistantly located 2 × 2 m quadrates along each transect. The coordinate of sampling points was registered for determining the elevation, slope and aspect using the topographic maps (Fig. 2). The maps of soil properties were produced interpolating the point data for each variable based on the kriging algorithm (Fig. 3). *Artemisia sieberi* was the most dominant species in the area due to its adaptation to a wide range of environmental conditions, and it was excluded from the analysis (Table 1). The vegetation types were modeled based on their second dominant species as indicator species where *A. sieberi* was the first dominant species.

BLR and CCA techniques: The presence/absence data for the plant species were used for CCA and BLR analysis. Binary logistic regression was calculated (Guisan *et al.*, 1999) as,

$$P = \frac{\exp(a + b_1x_1 + b_2x_2 + \dots + b_nx_n)}{1 + \exp(a + b_1x_1 + b_2x_2 + \dots + b_nx_n)}$$

This equation gives a probability value from 0 to 1, where p is the probability of the vegetation type occurrence, \exp is the base of natural logarithms, a is the constant of the equation, b is the coefficient of the explanatory variables and $x_1 \dots x_n$ are the variables. Backward stepwise logistic regression method was used to identify the statistically significant predictors ($P < 0.05$). The amount of variance explained by the predictors in each model was determined by calculating Nagelkerke's R square.

For modeling the species distribution using CCA, two matrices were constructed. In the main matrix, presence/absence data of the indicator species were used, and the second matrix was composed of the values of environmental variables (Ter Braak, 1986). The map of

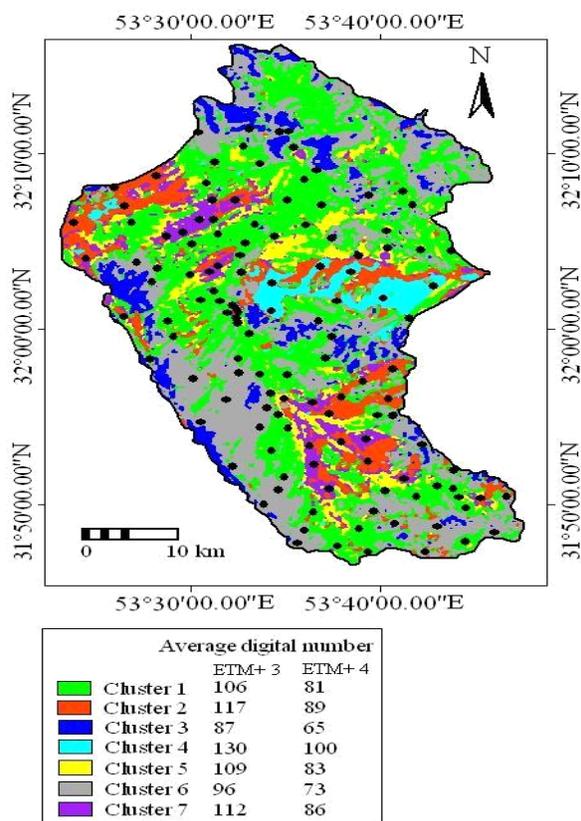


Fig 2. The classified map of Nodushan rangeland using unsupervised classification (clustering) algorithm based on the images of ETM+ bands 3 and 4. The sample points were selected based on overlaying the elevation, geologic and the cluster map. A 5×5 majority filter was applied to the clustered map.

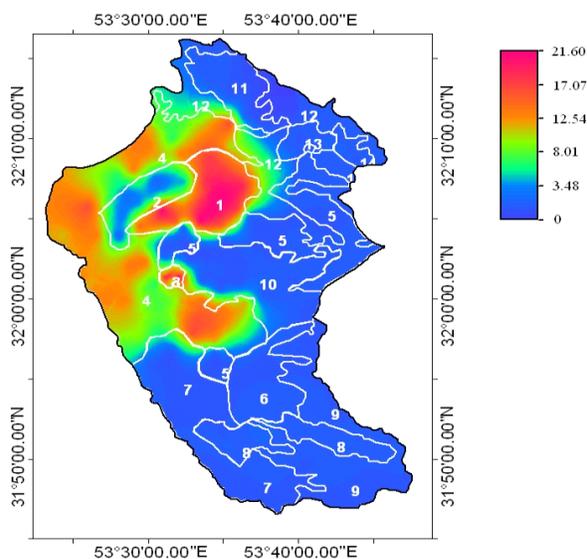


Fig 3. The map of EC (depth 0-20 cm) for Nodushan rangelands, an example of the produced map using kriging interpolation of point data

the environmental gradient for each canonical axis was produced for four first axes using the regression equation made with canonical coefficients. CCA was performed using Canoco 4.5 (ter Braak and Smlauer, 2002). The predicted probability of the species occurrence was classified based on the tolerance of the species. The tolerance based on presence-absence data is the standard deviation of the values of the ordination axis over the sampling points selected for each species (ter Braak, 1987). To identify the suitable = 1 and unsuitable = 0 habitat for each species based on CCA and BLR, the maximum sensitivity plus specificity threshold was calculated for each species model using the occurrence probability maps. This is one of the efficient methods for determining the threshold in species distribution modeling (Liu *et al.*, 2005). Sensitivity is calculated as $a/(a+c)$ and specificity as $d/(b+d)$, where a: true positives (or presences), b: false positives (or presences), c: false negatives (absences) and d: true negatives (or absences). Kappa coefficient was used to compare the predicted potential distribution maps with the actual vegetation map (Fig. 1) for obtaining the accuracy of the prediction. Kappa statistic strength of agreement is classified as: <0 poor, 0-0.20 slight, 0.21-0.40 fair, 0.41-0.60 moderate, 0.61-0.80 substantial and 0.81-1 almost perfect (Landis and Koch, 1977). The analysis and calculations were done using IIVIS 3.2 (ILWIS 3.2 Academic Version, 2004) and SPSS 15 (SPSS Inc., Chicago, IL).

Results and Discussion

BLR spatial modeling of the vegetation types: The significant predictors ($P < 0.05$), odds ratio, coefficient of variables and the value of Nagelkerke's R square were obtained from the BLR modeling of the vegetation types (Table 2). The strongest relationship between the prediction and the predictors was found for *I. songarica* (Nagelkerke $R^2 = 0.94$), followed by *E. strobilacea* (0.91), *A. aphylla* (0.90), *S. arbusculiformis* (0.89), *A. aucheri* (0.88), *S. tomentosa* (0.87), *E. ceratoides* (0.87) and *Z. eurypterum* (0.71). The lowest R-squared (0.59) was related to *P. harmala*. C/N_1 , $sand_1$, OM_2 and silt, were the significant predictors in predicting the occurrence of *I. songarica*. Gypsum_{1,2} made a significant contribution to the prediction of the presence probability of *E. strobilacea*. EC_{1,2} was the significant factor for *A. aphylla* and *S. tomentosa*, EC₁, C/N_1 and gypsum₂ for *S. strobilacea*, elevation for *A. aucheri*, C/N_1 for *E. ceratoides*, lime_{1,2} and $sand_2$ for *Z. eurypterum* and OM_1 , AW₂ and clay₂ for *P. harmala* model. Odds ratio (OR) is defined as the probability of occurrence over the probability of non-

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Table 1. Mean values of environmental variables and cover of the plant species (%) in nine studied vegetation types of Nodushan rangelands

	<i>A.sieberi- S. tomentosa</i>	<i>A.sieberi- E. strobilacea</i>	<i>A. aphylla- A. sieberi</i>	<i>A.sieberi-S. arbusculiformis</i>	<i>A.sieberi- P. harmala</i>	<i>A.sieberi- Z. eurypterum</i>	<i>A.sieberi-E. ceratoides</i>	<i>A.sieberi-I. songarica</i>	<i>A.aucheri</i>
Environmental variable									
Gravel ₁	17(4.58)	14(3.82)	12(3.97)	12(3.89)	11(4.69)	11(3.99)	16(3.86)	8(3.19)	8(3.9)
Gravel ₂	19(3.79)	21(3.73)	18(3.94)	21(1.99)	19(4.54)	14(4.78)	9(4.5)	14(4.53)	10(4.87)
PH ₁	7.3(0.24)	7.64(0.19)	7.43(0.23)	7.38(0.18)	7.59(0.19)	7.9(0.06)	7.06(0.07)	7.19(0.14)	7.67(0.13)
PH ₂	7.83(0.08)	7.31(0.12)	7.31(0.12)	7.43(0.13)	7.78(0.09)	7.75(0.1)	7.72(0.17)	7.16(0.09)	7.69(0.09)
Gyps ₁	0(0)	61.74(7.55)	2.06(0.73)	0(0)	0(0)	0(0)	1.01(0.21)	0(0)	0(0)
Gyps ₂	0(0)	63.43(6.43)	0(0)	29.4(2.64)	0(0)	0(0)	0.11(0.03)	0(0)	0.21(0.06)
C/N ₁	5.3(1.6)	3.94(1.71)	11.87(3.22)	1.38(0.55)	9.5(3.52)	11.75(4.56)	57.1(7.04)	6(1.89)	16.52(3.73)
C/N ₂	4.76(1.93)	75.71(4.81)	3.51(1.44)	2.76(0.97)	4.52(2.11)	4.06(2.09)	21.07(3.62)	3.45(1.05)	12.45(4.17)
Sand ₁	69.4(4.63)	66.4(3.94)	63.4(3.74)	57.4(4.03)	63.4(3.57)	53.4(2.89)	71.4(4.23)	43.4(3.83)	41.4(1.8)
Sand ₂	73.4(4.29)	81.4(3.01)	68.4(3.86)	83.4(4.92)	77.4(4.01)	70.4(4.27)	77.4(3.58)	63.4(3.2)	41.4(2.79)
Silt ₁	21.3(4.38)	23.3(3.78)	26.3(5.24)	33.3(4.18)	27.3(4.45)	37.3(4.72)	21.3(5.03)	47.3(5.3)	46.9(5.25)
Silt ₂	17.3(3.74)	11.3(3.57)	21.3(3.75)	11.3(2.56)	17.3(3.26)	22.3(3.88)	16.3(3.46)	27.3(3.39)	45.3(3.7)
Clay ₁	9.3(3.07)	10.3(2.72)	10.3(3.03)	9.3(2.62)	9.3(2.63)	9.3(2.91)	7.3(3.16)	9.3(3.56)	11.7(3.22)
Clay ₂	9.3(1.81)	7.3(2.14)	10.3(1.98)	5.3(2.09)	5.3(2.47)	7.3(1.96)	6.3(2.51)	9.3(2.47)	13.3(2.61)
AW ₁	0.1(0.02)	0.1(0.03)	0.11(0.02)	0.12(0.03)	0.11(0.02)	0.13(0.03)	0.09(0.03)	0.17(0.02)	0.17(0.02)
AW ₂	0.09(0.03)	0.05(0.01)	0.1(0.03)	0.03(0.01)	0.06(0.02)	0.09(0.01)	0.07(0.02)	0.11(0.03)	0.16(0.02)
EC ₁	16(3.14)	3.24(1.41)	16.5(3.77)	10.7(3.17)	0.673(0.24)	0.78(0.2)	0.72(0.24)	1.01(0.24)	0.624(0.24)
EC ₂	1.98(0.95)	6.01(1.86)	30.37(4.42)	5.3(1.71)	1.17(0.5)	1.047(0.35)	0.87(0.27)	0.52(0.25)	1.6(0.45)
Lime ₁	17.57(2.49)	9.07(3.27)	23.3(2.92)	22.8(4.06)	22.8(3.25)	34.8(3.52)	12.82(2.97)	34.8(2.5)	20.32(2.79)
Lime ₂	13.57(4.51)	9.07(3.83)	18.57(2.74)	13.5(4)	23.57(4.05)	43.3(4.84)	15.3(4.18)	50.32(3.58)	20.07(4.6)
OM ₁	0.075(0.01)	0.275(0.05)	0.229(0.05)	0.107(0.04)	0.458(0.12)	0.427(0.11)	0.275(0.05)	0.275(0.05)	0.917(0.16)
OM ₂	0.275(0.06)	0.091(0.03)	0.229(0.06)	0.153(0.05)	0.229(0.06)	0.382(0.04)	0.305(0.06)	0.382(0.05)	0.962(0.13)
Elevation	1850(18)	1940(22)	2010(11)	2074(31)	2186(54)	2262(31)	2318(29)	2420(42)	2618(46)

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Plant species	A.sieberi- S. tomentosa	A.sieberi- E. strobilacea	A. aphylla- A. sieberi	A.sieberi-S. arbusculiformis	A.sieberi- P. harmala	A.sieberi- Z. eurypterum	A.sieberi-E. ceratoides	A.sieberi-I. songarica	A.aucheri
Artemisia sieberi	6.6	5.1	2.9	9.0	5.2	12.9	10.4	16.6	-
Salsola tomentosa	1.5	0.1	-	0.15	-	-	-	-	-
Ephedra strobilacea	-	4.8	-	-	-	-	-	-	-
Anabasis aphylla	-	-	3.3	-	-	-	-	-	-
Salsola arbusculiformis	-	-	-	1.4	-	-	-	-	-
Peganum harmala	-	-	-	-	2.62	-	-	-	-
Zygophyllum eurypterum	-	-	-	-	-	2.6	-	-	-
Acantholimon sp.	-	-	-	-	-	1.9	-	2.1	-
Eurotia ceratoides	-	-	-	-	-	-	2.1	-	-
Astragalus glaucacanthus	-	-	-	-	-	-	0.52	-	-
Iris songarica	-	-	-	-	-	-	2.0	2.7	-
Stipa barbata	-	-	-	-	-	-	-	0.6	0.7
Artemisia aucheri	-	-	-	-	-	-	-	-	23
Astragalus micriphyssa	-	-	-	-	-	-	-	-	1.5
Scariola orientalis	-	-	-	0.1	-	-	-	-	-

Indices 1 and 2 indicate the first (0-20 cm) and the second (20-60 cm) soil layer, respectively; Organic matter was calculated by multiplying % organic carbon by 1.724; values in parenthesis indicate standard deviations.

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occurrence. The odds ratio of EC₁ (1.17) and EC₂ (1.48) for *A. aphylla* model indicated that for every one unit increase in EC₁ and EC₂ the odds of occurring this species increased by 17% and 48%, respectively. In addition the influence of EC₂ was greater than that of EC₁ on the occurrence probability of *A. aphylla*. The occurrence probability of *E. ceratoides* was positively associated with C/N₁ (OR: 1.22), and one unit increase in C/N₁ was associated with an increase of 1.22 times in the occurrence to non-occurrence probability of *E. ceratoides*. The odds ratio of elevation (1.02) in the *A. aucheri* model indicated that the probability of the presence to the absence of *A. aucheri* increased by 2% with an increase of one unit (meter) in the elevation. The calculated correlation coefficient (*r*) was 0.97 between the elevation and annual precipitation and -0.99 between the elevation and mean annual temperature in the study

area. This indicated that temperature and precipitation are the important drivers of *A. aucheri* distribution and that the elevation can be considered as an efficient indicator of changes in temperature and precipitation for modeling vegetation distribution in the study area. The odds ratio of gypsum₁ (2.46) and gypsum₂ (1.56) in the *E. strobilacea* model indicated that the influence of gypsum₁ was more than that of gypsum₂ on the probability of the occurrence of *E. strobilacea* and that the occurrence to non-occurrence probability of this species increased by a factor of 2.46 and 1.56 with one unit increase in gypsum₁ and gypsum₂, respectively. The odds of *Z. eurypterum* occurrence were positively associated with lime₁ (OR: 1.61), sand₂ (OR: 1.3) and lime₂ (OR: 1.1) and increased by 61, 30 and 10% as lime₁, sand₂ and lime₂ increased by one unit, respectively. The probability of occurrence to non-occurrence of *S.*

Table 2. Results of logistic regression using soil and topographic variables

Plant species	Nagelkerke R ²	Predictor	B	Sig.	Odds ratio
<i>Iris songarica</i>	0.94	C/N ₁	0.887	0.05	2.43
		Sand ₁	-1.11	0.04	0.33
		OM ₂	-0.40	0.044	0.67
		Silt ₁	0.41	0.038	1.50
		Constant	42.44	0.062	
<i>Ephedra strobilacea</i>	0.91	Gypsum ₁	0.9	0.05	2.46
		Gypsum ₂	0.468	0.044	1.56
		Constant	-5.06	0.001	
<i>Anabasis aphylla</i>	0.90	EC ₁	0.157	0.04	1.17
		EC ₂	0.389	0.03	1.48
		Constant	-7.135	0.004	
<i>Salsola arbusculiformis</i>	0.89	EC ₁	0.322	0.05	1.38
		C/N ₁	-0.55	0.034	0.58
		Gypsum ₂	0.07	0.04	1.07
<i>Artemisia aucheri</i>	0.88	E	9.98	0.43	
		Constant	0.018	0.002	1.02
		Constant	-46.37	0.002	
<i>Salsola tomentosa</i>	0.87	EC ₁	0.606	0.002	1.83
		EC ₂	0.011	0.04	1.01
		Constant	-5.65	0.037	
<i>Eurotia ceratoides</i>	0.87	C/N ₁	0.198	0.001	1.22
		Constant	-6.43	0.000	
<i>Zygophyllum eurypterum</i>	0.71	Lime ₁	0.476	0.013	1.61
		Lime ₂	0.093	0.045	1.10
		Sand ₂	0.267	0.014	1.30
		Constant	-38.96	0.004	
<i>Peganum harmala</i>	0.59	OM ₁	15.55	0.002	1.17
		AW ₂	-79.33	0.003	0.45
		Clay ₂	-0.45	0.050	0.64
		Constant	0.199	0.855	

Indices 1 and 2 indicate the first (0-20 cm) and the second (20-60 cm) soil layer, respectively

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tomentosa was increased by 83% and 1% with one unit increase in EC_1 and EC_2 , respectively. The probability of occurrence of *S. arbusculiformis* was positively associated with EC_1 (OR: 1.38) and gypsum₂ (OR: 1.07) and negatively associated with C/N_1 (0.58). In this case, the odds of *S. arbusculiformis* presence increased by 1.38 and 1.07 times with one unit increase in EC_1 and gypsum₂, respectively and decreased by 0.58 times with one unit increase in C/N_1 .

Table 3. Kappa coefficient for the predicted distribution of indicator plant species using BLR and CCA

Plant species	Kappa BLR model	Kappa CCA model
<i>Iris songarica</i>	0.65	0.40
<i>Ephedra strobilacea</i>	0.79	0.75
<i>Anabasis aphylla</i>	0.67	0.79
<i>Salsola arbusculiformis</i>	0.78	0.85
<i>Artemisia aucheri</i>	0.71	0.76
<i>Salsola tomentosa</i>	0.70	0.68
<i>Eurotia ceratoides</i>	0.90	0.54
<i>Zygophyllum eurypterum</i>	0.72	0.52
<i>Peganum harmala</i>	0.37	0.04

C/N_1 had the greatest influence on the occurrence probability of *I. songarica* followed by sand₁, silt₁ and OM₂. The odds of occurring *I. songarica* increased by a factor of 2.43 and 1.5 with one unit increase in C/N_1 and silt₁ and decreased by a factor of 0.33 and 0.67 with one unit increase in sand₁ and OM₂. The probability of presence to absence of *P. harmala* increased by 1.17-fold and decreased by 0.45 and 0.64-fold with one unit increase in OM₁, AW₂ and clay₂, respectively. The highest accuracy of prediction (Table 3) was found for *E. ceratoides* (kappa = 0.90) with an almost perfect agreement between the predicted distribution map and the actual vegetation map followed by *E. strobilacea* (0.79), *S. arbusculiformis* (0.78), *Z. eurypterum* (0.72), *A. aucheri* (0.71), *S. tomentosa* (0.7), *A. aphylla* (0.67) and *I. songarica* (0.65) with a substantial accordance between the predicted and the actual distribution. The lowest accuracy was related to the prediction of *P. harmala* occurrence (0.37) with a fair agreement.

CCA spatial modeling of the vegetation types: The ordination biplots for the nine indicator plant species showed the overall results of the CCA ordination. The CCA explained 49.2% of the variance in species presence-absence data. Axis 1, 2, 3 and 4 explain 12.5%, 12.3%, 12.2% and 12.2% of the variability in the species data, respectively. Of the 25 variables, 20 variables had the highest correlation with the first two axes. pH₂, gypsum

C/N_2 and lime_{1,2} had the highest correlation with the first axes. High score of *E. strobilacea* on axis 1 indicated that this species is found on the soils with a high gypsum_{1,2} and C/N_2 and a low lime_{1,2} and pH₂ (Fig. 4). Gypsum_{1,2} and C/N_2 appear to have more influence on the distribution of this species than lime_{1,2} and pH₂ based on their correlation to axis 1. The relatively low scores of *Z. eurypterum* and *I. songarica* on axis 1 exhibited a preference for high-lime soils. Gravel_{1,2}, pH₁, sand_{1,2}, silt_{1,2}, EC₁, OM_{1,2}, AW_{1,2}, E and slope had the highest correlation with the second axes. The species were sorted out well along this axis. High score of *A. aucheri* on axis 2 indicated that this species is a high elevation and steep slope species and showed a preference for silty loam soils with high OM_{1,2} and AW_{1,2} and low gravel_{1,2} and salinity₁. *A. aphylla*, *S. tomentosa* and *S. arbusculiformis* on axis 2 indicated that these species are low elevation species and had a preference for sandy loam soils with high salinity₁ and gravel_{1,2}. Although EC₂ had the highest correlation with axis 4 but its high correlation to axis 2 indicated that this factor is also an important driver in the occurrence of *A. aphylla*, *S. tomentosa* and *S. arbusculiformis*. *I. songarica* on axis 1 and 2 indicated that this species is a relatively high elevation species with a preference for silty loam soils. *P. harmala*, *E. ceratoides* and *Z. eurypterum* were dominant at middle elevation. Axis 3 was dominated by C/N_1 . *S. arbusculiformis* was sorted out well along this axis indicating that this species exhibited a preference for a low C/N_1 . Axis 4 was dominated by clay_{1,2} and EC₂. In addition, EC₁ had also a high correlation with this axis. High score of *A. aphylla* on axis 4 indicated that this species occurred on saline soils and preferred a heavier textured soil (Fig. 5). However, EC was appeared to be the most important driver of *A. aphylla* distribution. High score of *E. ceratoides* on axis 3 and its low score on axis 4 showed that the preference of this species for growing was the loamy sand soils with a high C/N_1 and low salinity. Low score of *I. songarica*, *Z. eurypterum* and *P. harmala* on axis 4 showed that these species were often found on non-saline soils. The highest accuracy of prediction was related to *S. arbusculiformis* with an almost perfect accordance (kappa = 0.85) between the predicted and actual distribution (Table 3). There was a substantial agreement for *A. aphylla* (0.79), *A. aucheri* (0.76), *E. strobilacea* (0.75) and *S. tomentosa* (0.68) model, a moderate agreement for *E. ceratoides* (0.54) and *Z. eurypterum* (0.52) and a fair agreement for *I. songarica* (0.40) model. The poorest accuracy of prediction was found for *P. harmala* (0.04). Overall, the prediction accuracy for *S. tomentosa*, *Z. eurypterum*, *E.*

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ceratoides, *E. strobilacea* and *I. songarica* distribution was higher by BLR than by CCA, whereas the distribution of *A. aphylla*, *S. arbusculiformis* and *A. aucheri* was more accurately predicted by CCA. The only exception was *P. harmala* with the poorest accuracy of prediction by both BLR and CCA. However, BLR provided more accurate prediction than CCA for this species. *P. harmala* appeared following the omission of plant species under an intense grazing pressure by sheep mainly in the areas close to villages in Nodushan rangelands. This was in line with the statement of Guisan *et al.* (1999) that the species for which model predictions had the poorest accuracy were under disturbance. The soil factors were the most important drivers of the distribution and occurrence of 8 vegetation types in arid rangelands except *A. aucheri* type, a semiarid vegetation type, whose distribution was affected by elevation. Elevation (Hosseini *et al.*, 2013) and precipitation (Yaghmaei *et al.*, 2008) were addressed as important factors influencing the distribution of *A. aucheri* in a local scale. Although studies showed that the species, whose distribution was well modeled by BLR, was also well modeled by CCA (Guisan *et al.*, 1999; Chahouki *et al.*, 2010), and vice versa. In the present study the vegetation types whose distribution was well modeled by BLR was not well modeled by CCA. BLR provided a perfect agreement for 1 species, a substantial agreement for 7 species and a fair agreement for 1 species, whereas CCA provided a prediction with a perfect agreement for 1 species, a substantial agreement for 4, a moderate for 2 and a fair and slight for 1 species. It was observed that BLR might give better species-specific models than CCA. Presence-absence data of species were found to be effective for an accurate modeling of vegetation distribution. This makes it needless to measure time-consuming quantitative attributes of plants such as density and cover. Determination of species-environment relations using CCA enabled us to simultaneously assess the influence of a variety of factors on the distribution of the species. Although the cumulative proportion of the variance of the species data was low, the predictions by CCA model were reasonable. Studies showed that an ordination axes that explained only a low amount of variance may still be quite informative (Gauch, 1982).

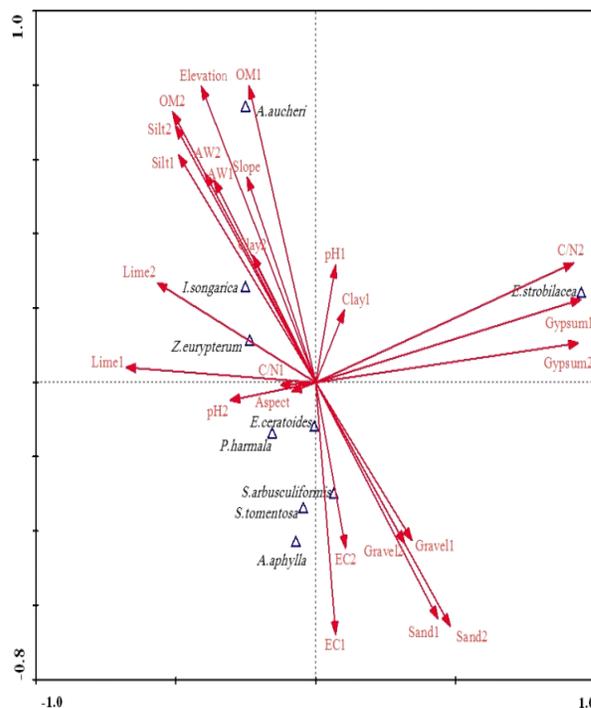


Fig 4. CCA biplots of the indicator plant species based on axis 1 versus axis 2. Species scores are linear combination scores

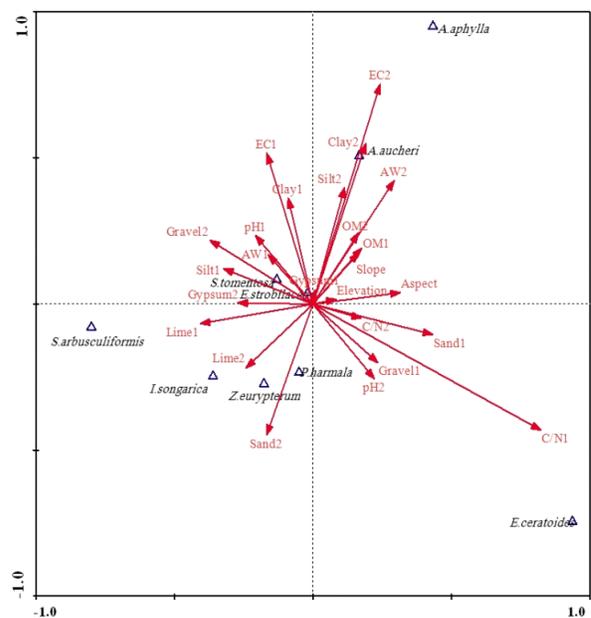


Fig 5. CCA biplots of the indicator plant species based on axis 3 versus axis 4. Species scores are linear combination scores

Conclusion

It was concluded that BLR models gave better predictions than CCA models for most of the study vegetation types because a species-specific subset of environmental predictors can be selected in BLR. Although a number of vegetation types were better modeled by CCA than by BLR, these vegetation types were well modeled by BLR too. Both BLR and CCA have limitations for modeling distribution of the vegetation types that are related to the grazing disturbance. However in such case, BLR is preferred over CCA. It is recommended that the BLR be used for modeling the distribution of vegetation types based on indicator species in arid and semiarid climates. Accurately prediction of species distribution enables an effective degraded land restoration planning and rangelands rehabilitation measures.

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